

Critical Loads for the Management of Nitrogen Acidification and Eutrophication

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Overview

- 1. The critical loads concept
- 2. Critical loads for N and acidity
- 3. Critical loads for N as a nutrient
- 4. Dynamic models
- 5. Critical loads in Alberta
- 6. Conclusions



1. The Critical Loads concept



Eagle Mountains, Czech Republic, 2005



Regulating long-range pollutant emissions

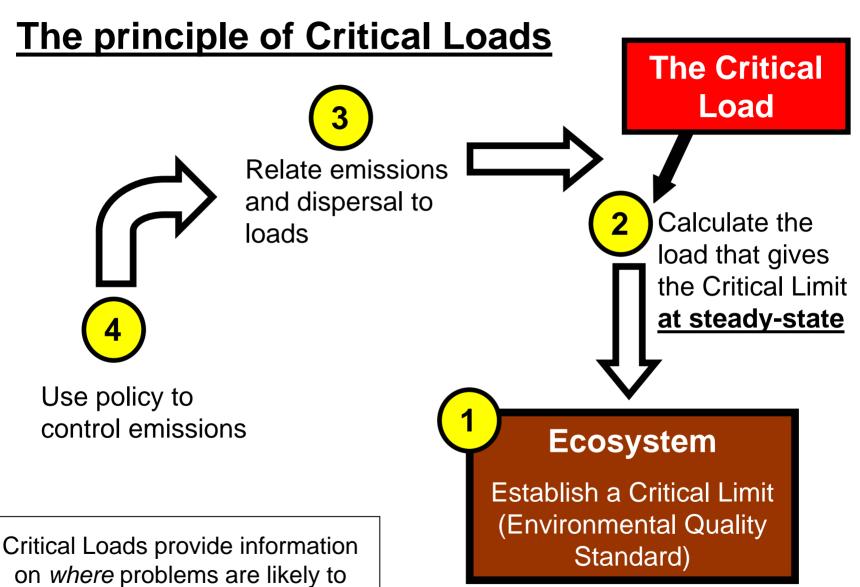
Option 1: Best-available technology

Option 2: Effects-based approach

<u>Critical Load</u> = the highest annual input of the pollutant that, at steady-state, does not cause unacceptable ecological [or human health] effects

<u>Critical Limit</u> = the highest steady-state concentration of the pollutant that does not cause unacceptable ecological [or human health] effects





occur

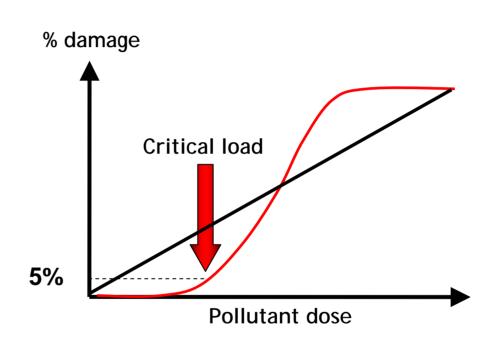


Setting critical loads

1. Define an indicator of change for the receptor of interest:

Ecosystem structure
Sensitive indicator species
Nitrate leaching
Soil acidification

- 2. Define a dose-response function
- 3. Define a damage threshold for the required level of ecosystem protection



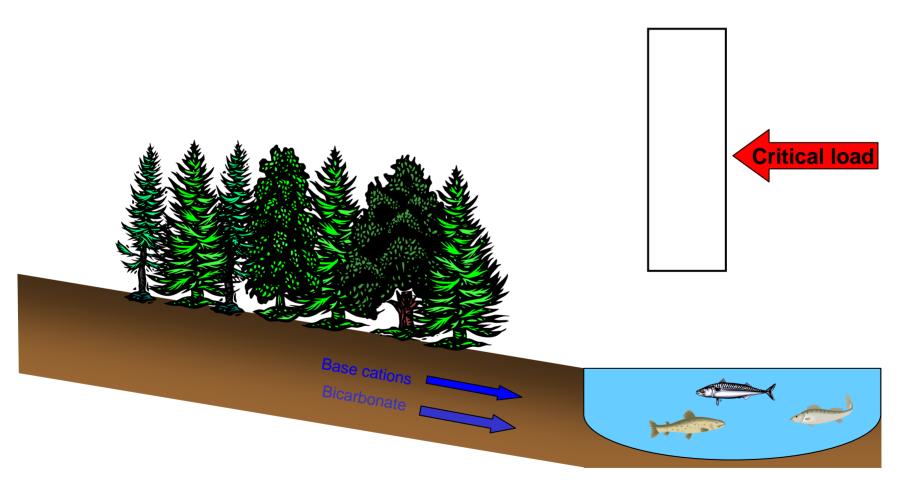
CLs assume a damage threshold exists – if dose vs damage is linear, we have more of a problem...



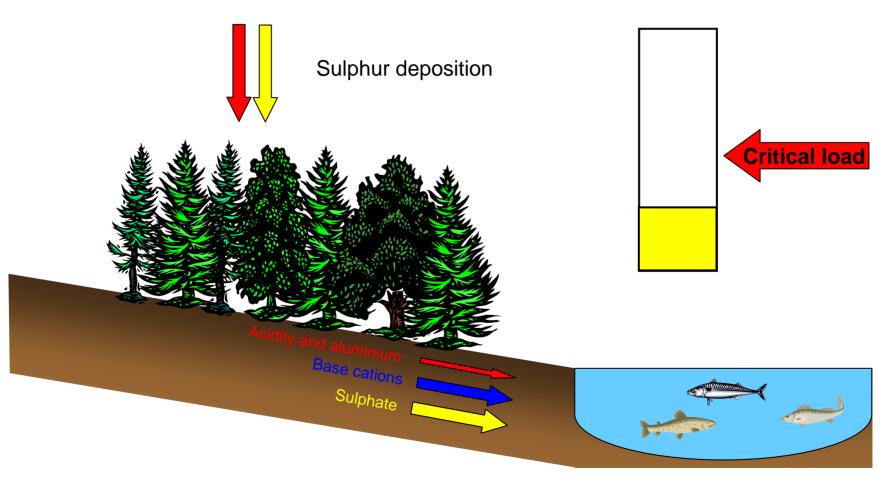
2. Critical Loads for acidification



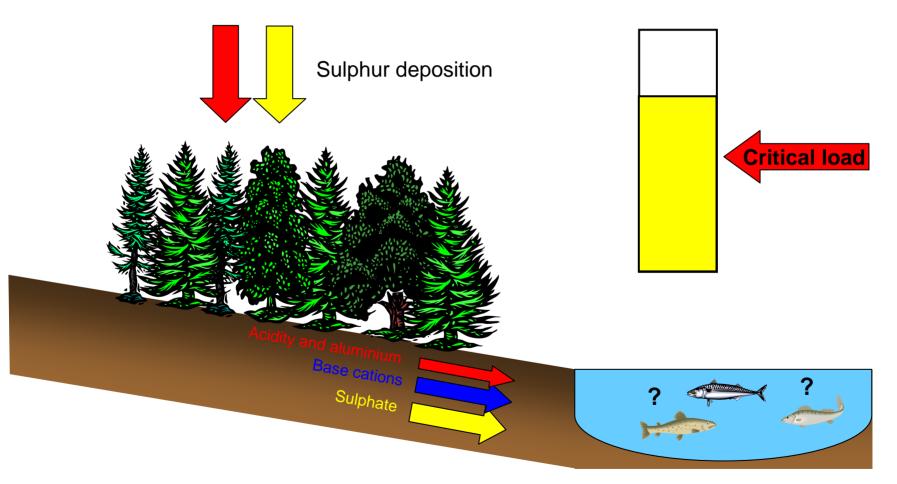




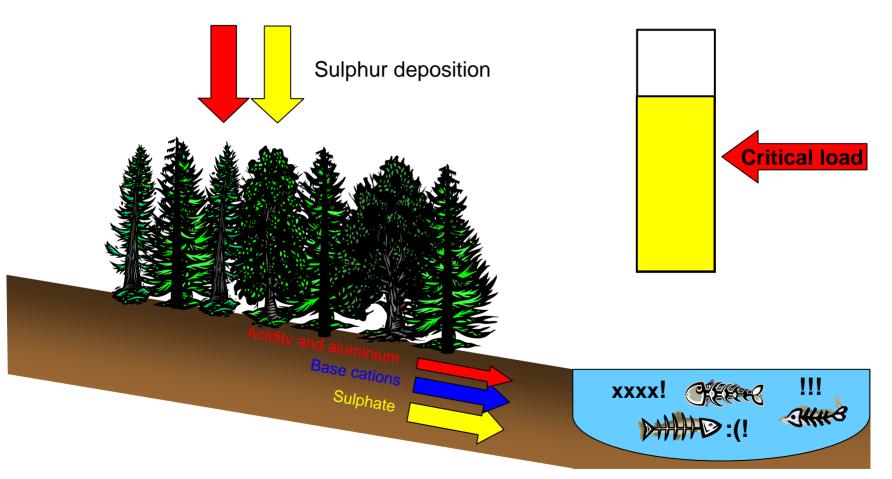






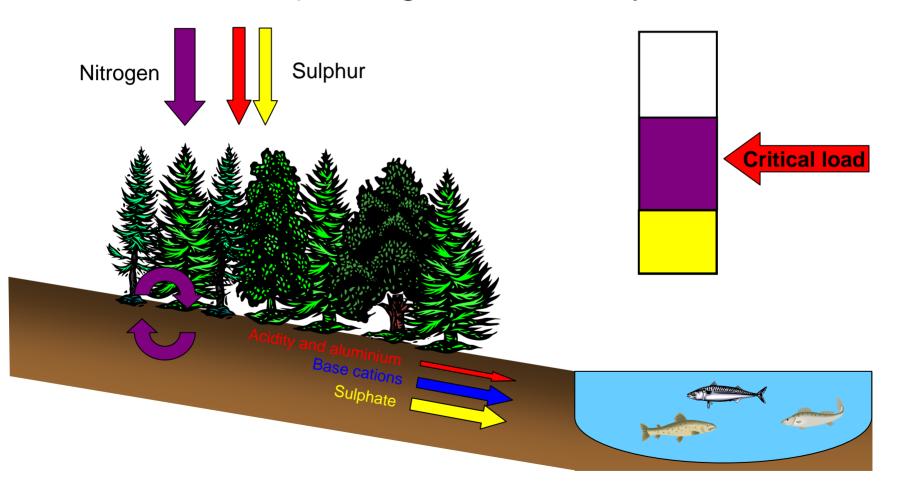






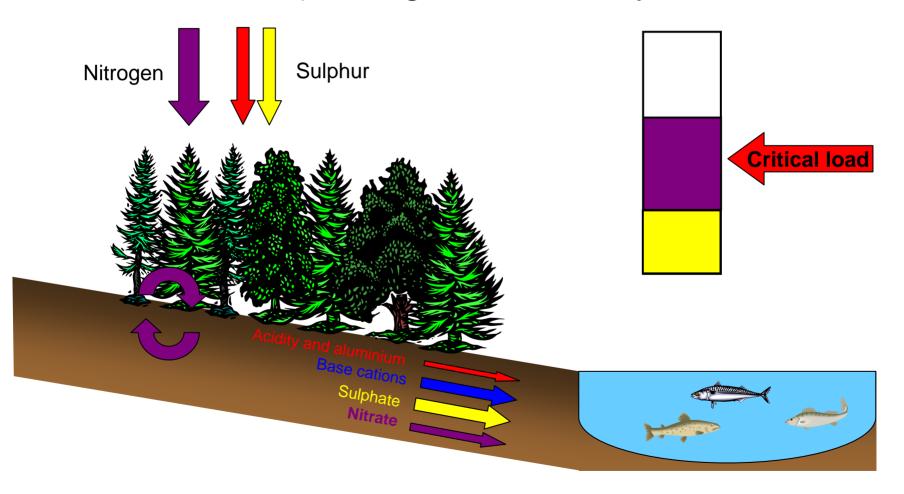


Critical Loads and ecosystem damage 1) Nitrogen and acidity



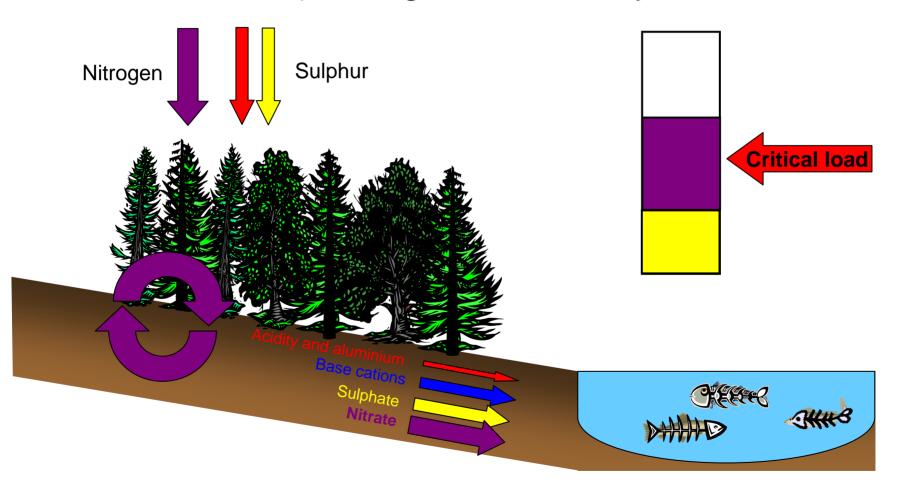


Critical Loads and ecosystem damage 1) Nitrogen and acidity





Critical Loads and ecosystem damage 1) Nitrogen and acidity





Calculating critical loads for acidity UK Methods

- Skokloster classes
 - Heathland and grassland
 - Basically estimates of long-term buffering provided by weathering in different soils
 - 5 sensitivity classes
- Simple mass balance (SMB)
 - Forests

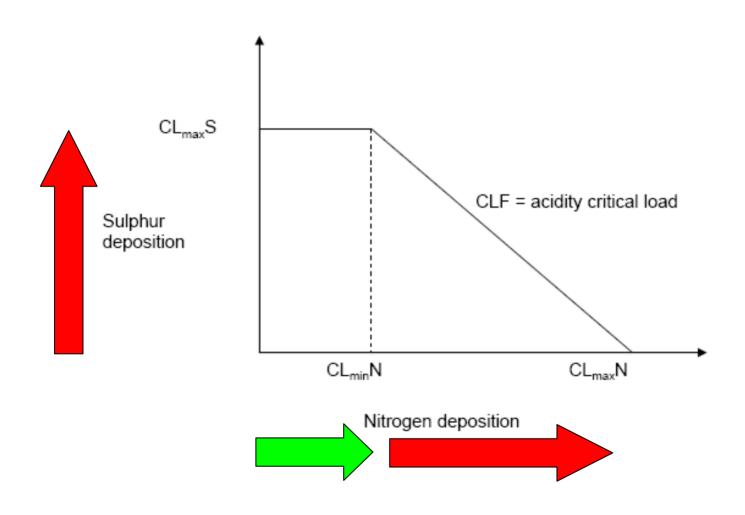
Simple mass balance (SMB) model

- Based on a critical limit for UK forests this is Ca:Al = 1
- Balances acid inputs and outputs to derive a critical load that ensures the critical limit is not exceeded
- And the equations are...

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\begin{split} &CL_{max}(S) = BC_{dep} - CI_{dep} + BC_w - BC_u + (1.5 \times Ca_{le}/(Ca:AI)_{crit}) + Q^{2/3}(1.5 \times Ca_{le}/((Ca:AI)_{crit} \times K_{Gibb})) \\ &CL_{min}(N) = N_i + N_{de} + N_u \\ &CL_{max}(N) = CL_{max}(S) + CL_{min}(N) \end{split}
```

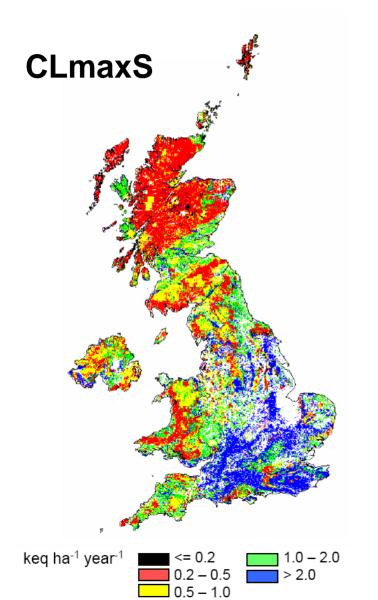


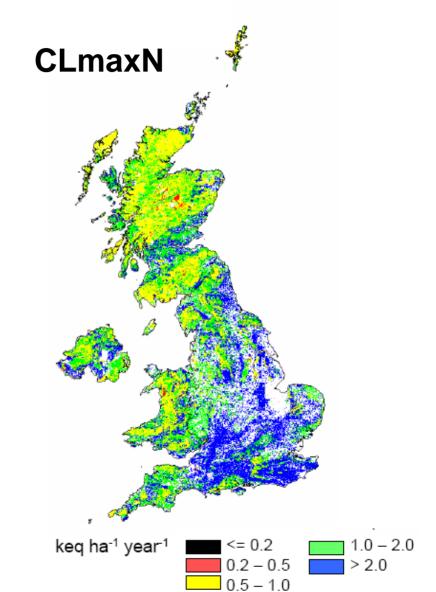
The critical load function





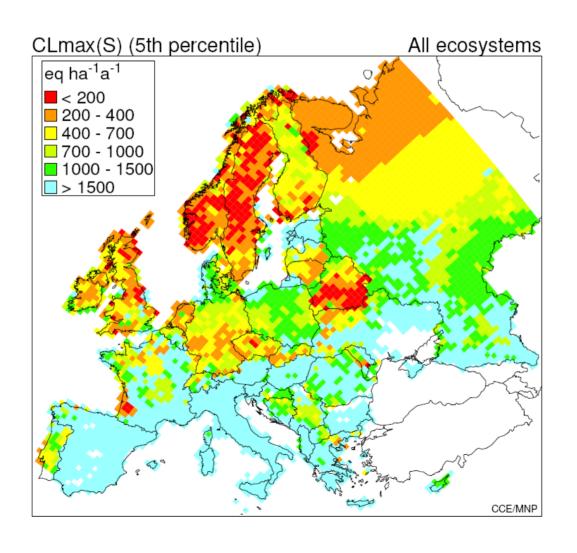
UK 5th percentile Critical Loads for Acidity



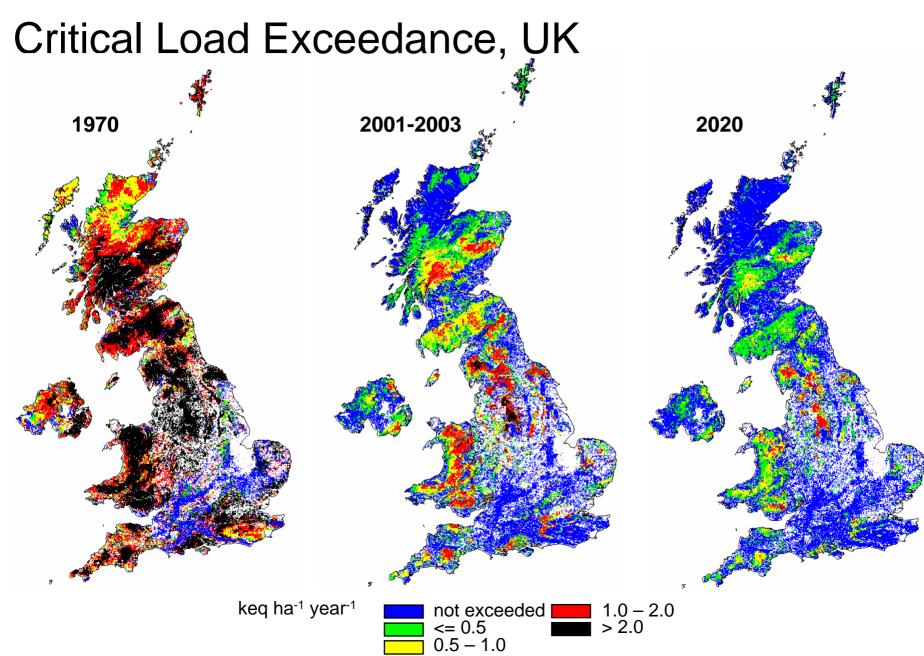




European Critical Loads for Acidity

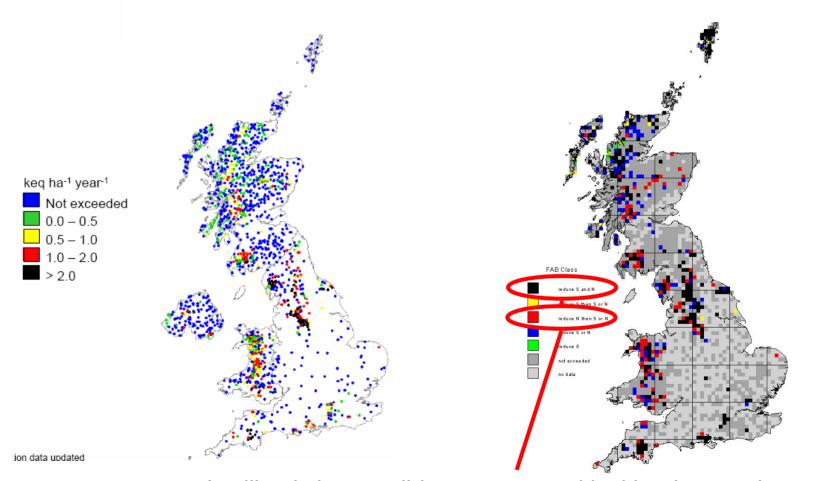








Critical Load exceedance in UK surface waters: What needs to be done to reduce exceedance?



It will only be possible to remove critical load exceedance in these areas by reducing N deposition

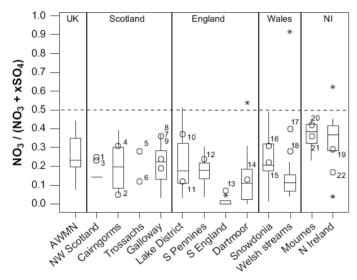


Time lags between exceedance and damage

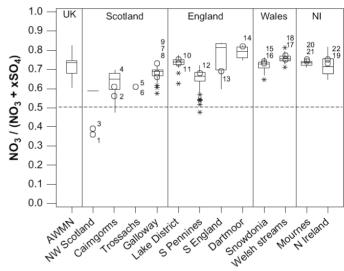
- For S deposition, exceedance of critical loads may lead to relatively rapid damage
- Delays occur due to:
 - Base cation buffering
 - S adsorption (mainly in unglaciated soils)
 - S reduction (mainly in wetlands)
- For N deposition, lags between critical load exceedance and damage may be much longer.
- Delays are primarily due to soil N immobilisation



Now



Predicted steady state (given 2010 deposition)



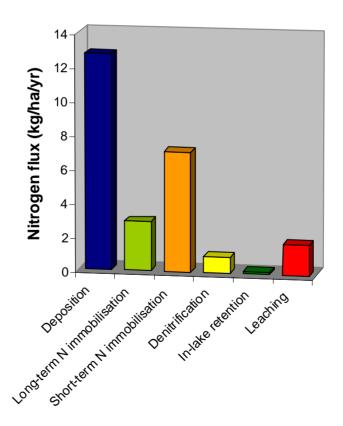
Curtis et al., Environmental Pollution (2005)

- Critical loads models generally predict a much higher level of steadystate N leaching than is currently observed
- Lag times appear to be long
- But if NO₃ leaching does reach predicted levels, future acidification could be as bad, or worse, than the 1970s-80s.



Nitrogen sources and sinks at Llyn Llagi, Wales

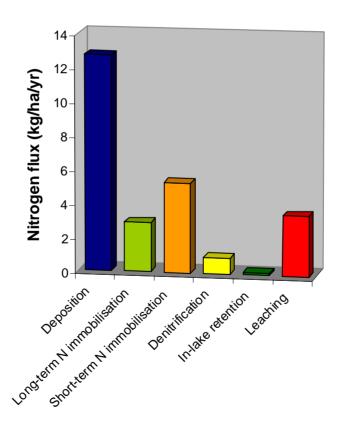
1. Present day





Nitrogen sources and sinks at Llyn Llagi, Wales

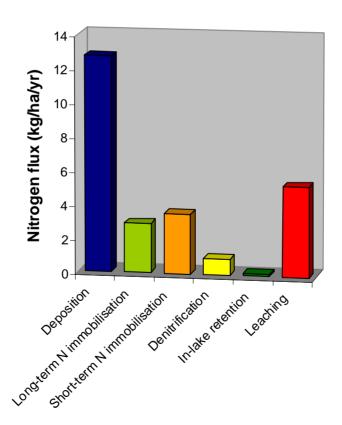
2. Future (1)





Nitrogen sources and sinks at Llyn Llagi, Wales

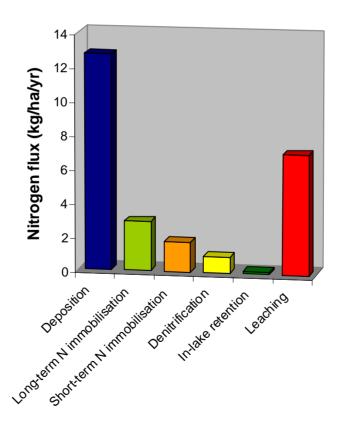
2. Future (2)





Nitrogen sources and sinks at Llyn Llagi, Wales

2. Future (3)



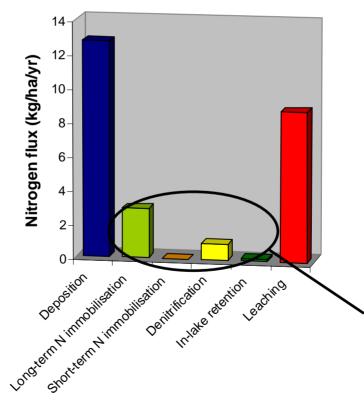


Nitrogen sources and sinks at Llyn Llagi, Wales

3. Steady State

But not all sites like this – some parts of Europe leaching most or all of incoming N already

In a managed forest, N uptake may reduce N leaching (but, N deposition may be higher)



Critical load for N only considers the sinks still operating at steady state



3. Critical Loads for Nitrogen as a Nutrient





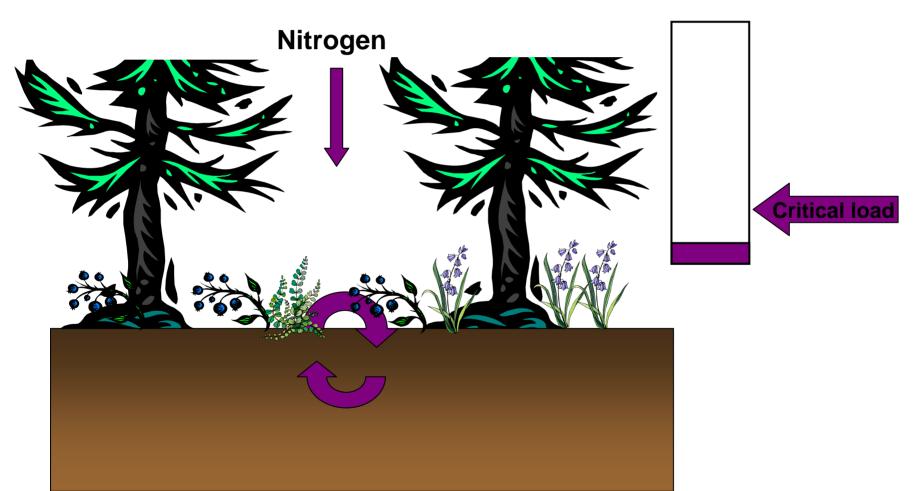
Nitrogen as a nutrient

- Nitrogen is a major nutrient required by all plants, and the limiting nutrient in most northern ecosystems
- Many natural habitats are characterised by slowgrowing species adapted for low-N conditions.
- With increased N deposition, these species are out-competed by faster-growing species more able to exploit increased N availability
- The results is a loss of biodiversity, or of characteristic plant species.



Critical Loads and ecosystem damage

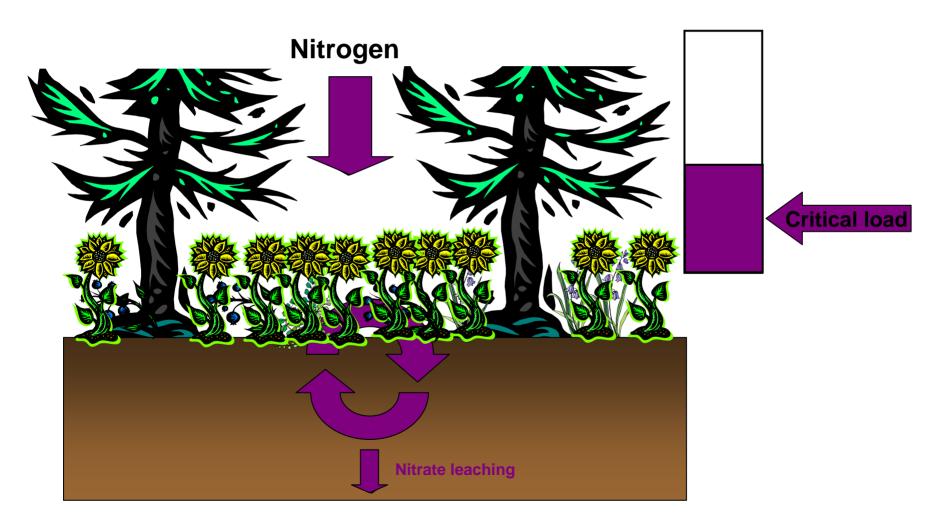
3) Nitrogen and biodiversity





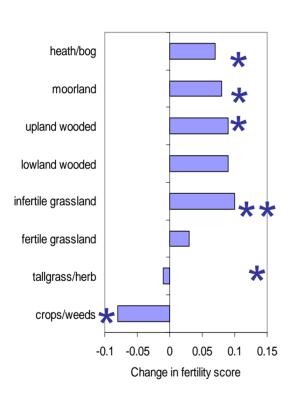
Critical Loads and ecosystem damage

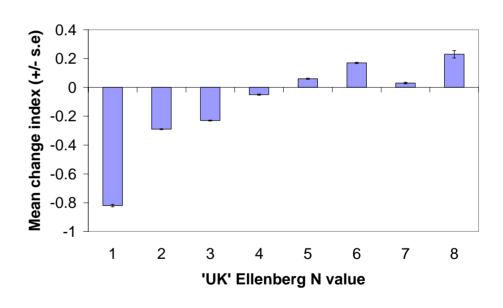
3) Nitrogen and biodiversity





Evidence that N deposition is causing eutrophication of UK ecosystems





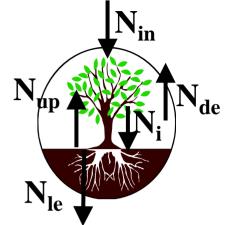
Countryside Survey, changes between 1990 and 1998

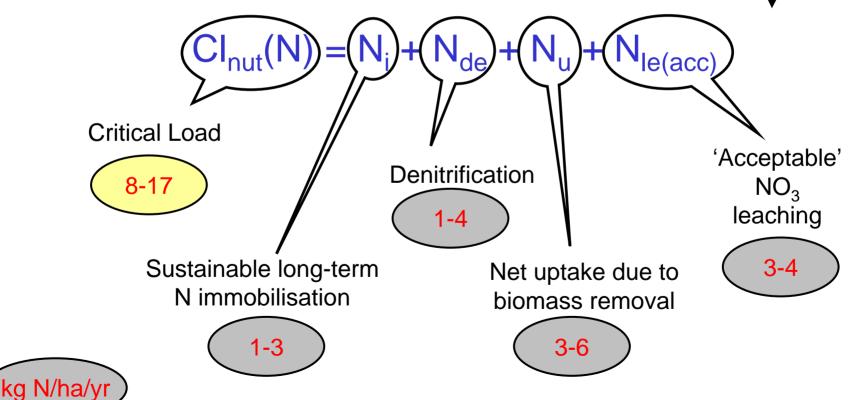
Plant Atlas, changes between 1930-69 and 1987-99



Critical Loads for N as a nutrient

Mass balance equation (used for UK managed forests)







Critical Loads for Nitrogen as a nutrient

Empirical critical loads (Used for other UK ecosystems)

- Based on experimental/field evidence of thresholds for change in species composition, plant vitality or soil processes
- Focused on communities likely to be sensitive to N deposition, of conservation value and with a reasonably wide distribution
- European ranges defined at a workshop in Berne, 2002
- Reliant on a large amount of scientific data, and a certain amount of expert judgement
- Countries decide which communities to protect, and where within the range to set their critical loads



Berne empirical critical loads, and their application in the UK

| (a) Ecosystem (with corresponding EUNIS class, where used) | (b) 2001 UK mapping value | (c) Critical load range in 1996 Mapping Manual | (d) Critical load range from Berne workshop | (e) Revised UK mapping value |
|--|-------------------------------|---|---|---|
| Grasslands Dry acid and neutral closed grassland (E1.7) Calcareous grassland (E1.26) Montane grassland Hay meadows (E2.2) Montane hay meadows (E2.3) Arctic/sub-alpine grass Moist/wet oligotrophic grass (E3.5) Molinia meadows (E3.51) Nardus stricta swards (E3.52) Moss/lichen mountain summits (E4.2) Inland dune pioneer grass (E1.94) Inland dune silicaceous grass (E1.95) | 25 25 ⁽¹⁾ 12 | 20-30 # 15-35 # 10-15 (#) | 10-20 # 15-25 ## 20-30 (#) 10-20 (#) 10-15 (#) 10-20 # 15-25 {#) 10-20 # 5-10 # 10-20 (#) 10-20 (#) | 15 20 - - - 15 - 15 7 |
| Heathland/moorland Lowland dry heaths (F4.2) Lowland Erica wet heaths (F4.11) Upland Calluna wet heaths (F4.11) Arctic/alpine heaths (F2) Tundra (F1) | 17 15 7.5 | 15-20 ## 17-22 # 10-20 (#) 5-15 (#) | 10-20 ## 10-25 # 10-20 (#) 5-15 (#) 5-10 # | 12 15 15 - |
| Coastal habitats Coastal stable dune grasslands (B1.4) Shifting coastal dunes (B1.3) Coastal dune heaths (B1.5) Moist-wet dune slacks (B1.8) Dune slack pools (C1.16) Salt marshes (A2.64 & A2.65) | | 20-30 # | 10-20 # 10-20 # 10-20 (#) 10-25 (#) 10-20 (#) 30-40 (#) | 15 15 - - - |
| Softwater oligotophic lakes Permanent oligotrophic lakes (C1.1) Bogs, mires and fens Ombrotrophic and raised bogs (D1) Poor fens (D2.2) Rich fens (D4.1) Montane rich fens (D4.2) | 10 | 5-10 ## 5-10 # | 5-10 ## 5-10 ## 10-20 # 15-25 (#) 15-25 (#) | - 10 15 - |



Examples of evidence underpinning Berne empirical CLs

- Boreal forest
 - 10-20 kg N/ha/yr, 'quite reliable'
- Onset of NO₃ leaching, N mineralisation
 - forest surveys, fertilisation experiments
- N/P and N/Mg imbalances in trees
 - forest surveys, fertilisation experiments
- Ground vegetation change
 - fertilisation experiments (e.g. displacement of Vaccinium myrtillus by Deschampsia flexuosa at > 5 kg N/ha/yr in N. Sweden)



Examples of evidence underpinning Berne empirical CLs

2) Tundra

5-10 kg N/ha/yr, 'quite reliable'

- Vegetation change
 - One set of fertilisation experiments receiving 10 kg N/ha/yr, Svalbard, showing changes in species composition of moss layer, decrease in lichens.



Examples of evidence underpinning Berne empirical CLs

- 3) Alpine grasslands
 - 10-15 kg N/ha/yr, 'expert judgement'
- Vegetation change
 - One experiment in Switzerland showing biomass increase after 4 years addition of 20 kg N/ha/yr
- Extrapolation from (better studied) lowland grasslands

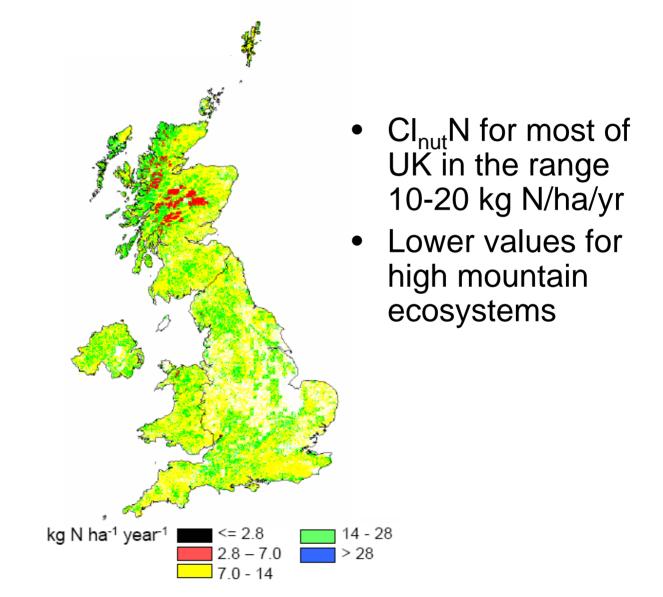


Examples of evidence underpinning Berne empirical CLs

- 4) Blanket bogs
 - 5-10 kg N/ha/yr, 'reliable'
- Increased N in peat and peat water
 - Experiments, field surveys
- Changes in moss growth and N content
 - Experiments, field surveys
- Increases in vascular plants over mosses
 - Experiments, field surveys

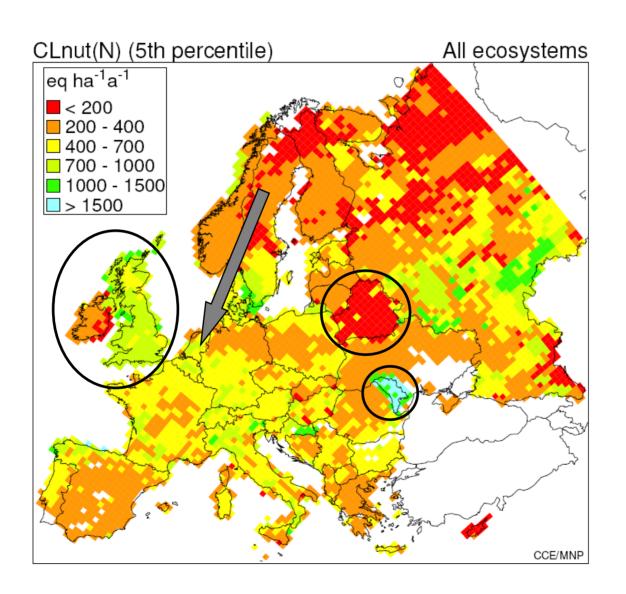


UK 5th percentile nutrient N critical loads



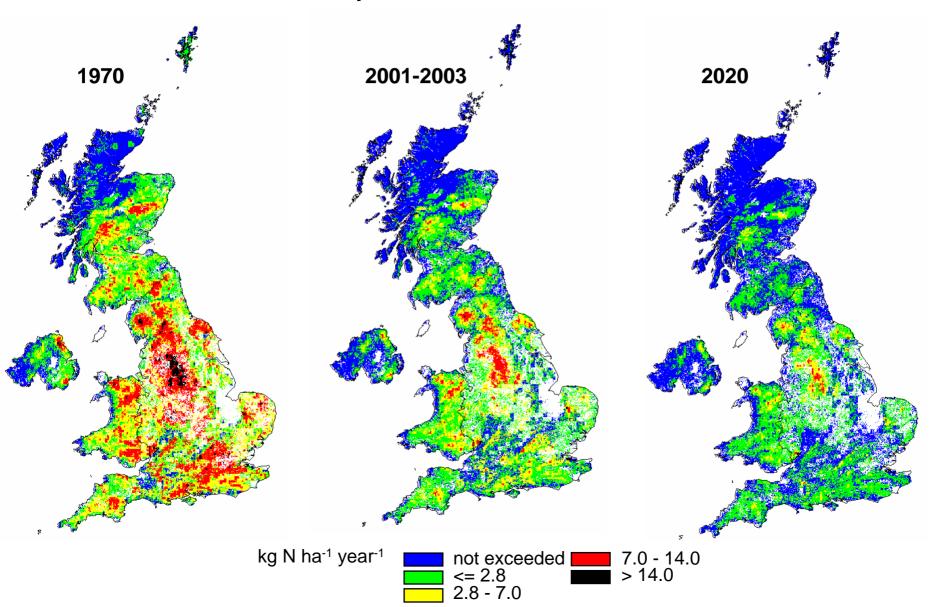


European 5th percentile nutrient N critical loads



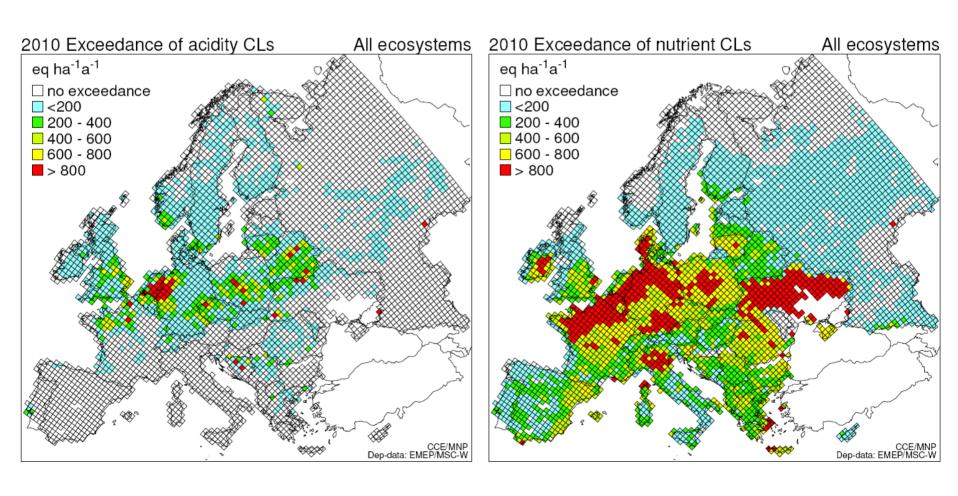


Exceedance of 5th percentile nutrient N critical loads



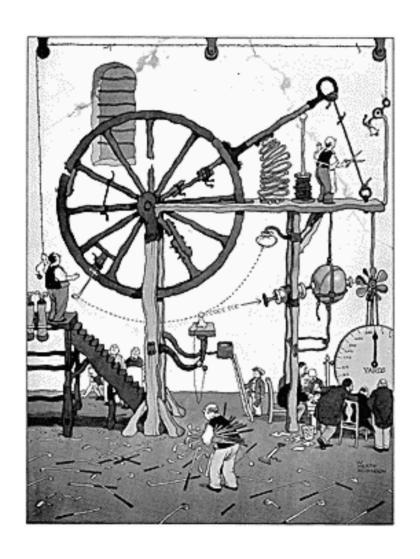


Critical load exceedances across Europe 2010 forecast





4. Dynamic Models



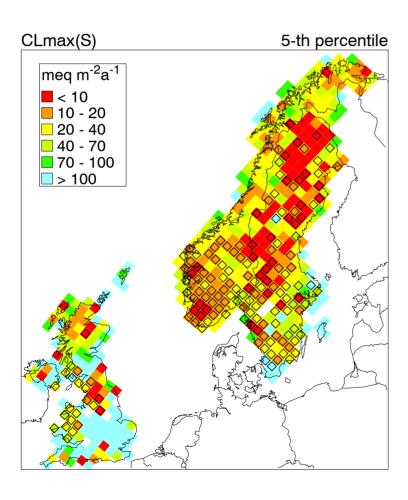


Dynamic Models

- Critical loads are essentially models of steadystate chemistry
- Dynamic models predict the time at which damage (or recovery) will occur
- Much current work in Europe is focused on modelling, in particular:
 - Setting 'Target Loads' the target deposition required to achieve acceptable chemical status by a given target date
 - Modelling biodiversity impacts by relating vegetation status to soil chemical status



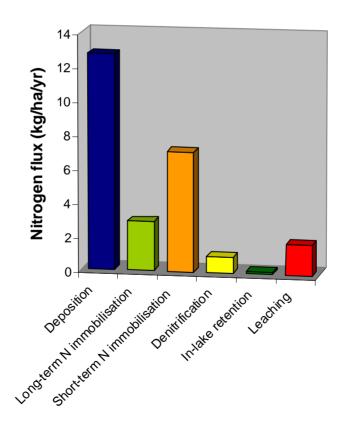
Target loads for acidity, N European surface waters





Modelled lags in N leaching (Llyn Llagi again)

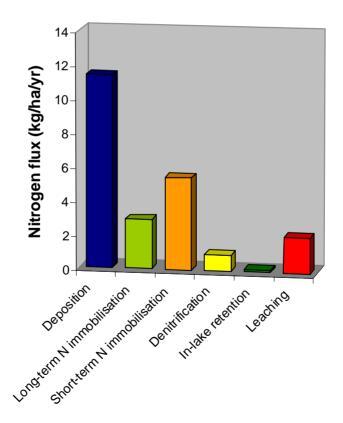
MAGIC calibrated present day





Modelled lags in N leaching (Llyn Llagi again)

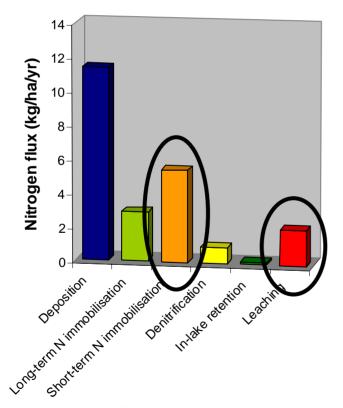
MAGIC predicted 2100



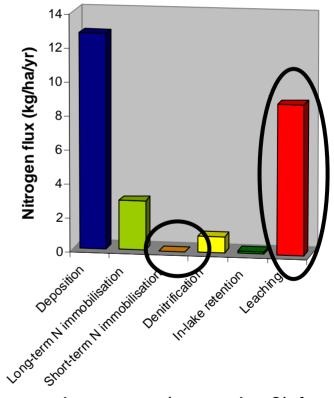


MAGIC modelling of lags in N leaching (Llyn Llagi again)

MAGIC predicted 2100



Predicted Steady State

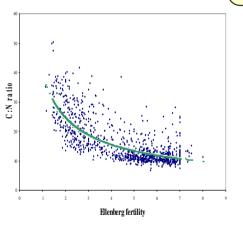


Dynamic models suggest that many ecosystems are a long way (centuries?) from the steady state NO₃ leaching levels indicated by the steady state mass balance



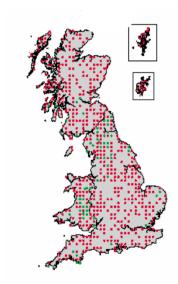
Predicting biodiversity change with dynamic models: MAGIC-GBMOVE

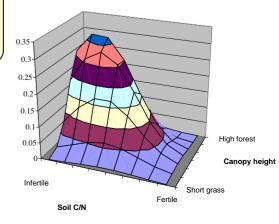
UK Countryside Survey: 16,691 vegetation survey plots. Species recorded, Ellenberg values for fertility (Eb N), acidity (Eb R) and moisture (Eb F) calculated



Subset of sites to relate Ellenberg values to abiotic conditions (soil pH, moisture, C/N ratio)

GBMOVE: Empirical relationships derived to predict probability of occurrence as a function of nitrogen, acidity and other environmental drivers







Predicting biodiversity change with dynamic models: MAGIC-GBMOVE

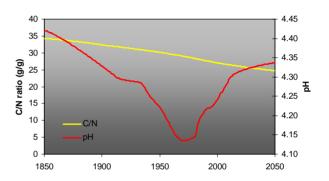
MAGIC: Prediction of soil pH and C/N change in response to changing S and N deposition Other environmental data

e.g. moisture, temperature, grazing

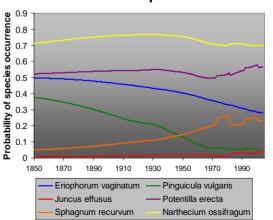
GBMOVE: Empirical relationships derived to predict probability of occurrence as a function of nitrogen, acidity and other environmental drivers

Note that GBMOVE does not assume a threshold

MAGIC simulation

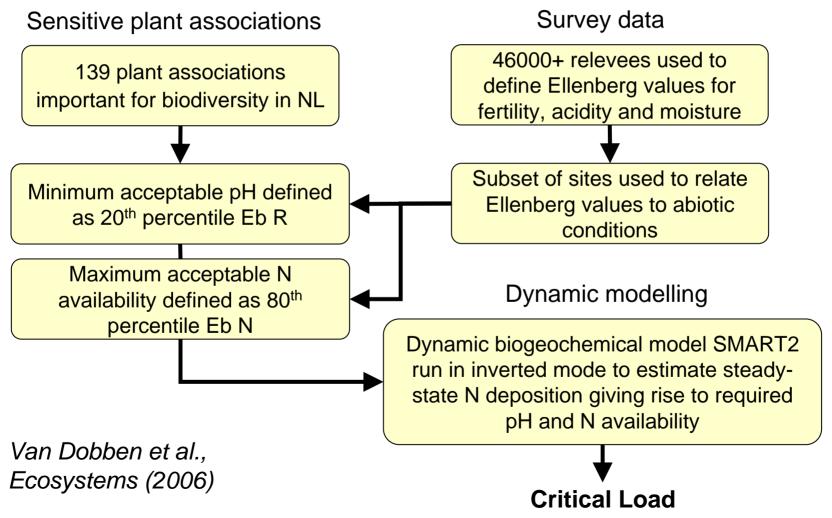


GBMOVE predictions



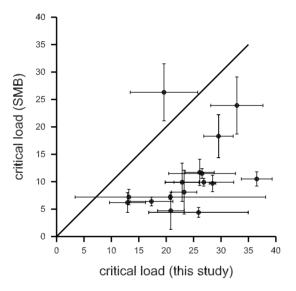


Calculating critical loads with dynamic models: 1. Netherlands





Calculating critical loads with dynamic models: 1. Netherlands



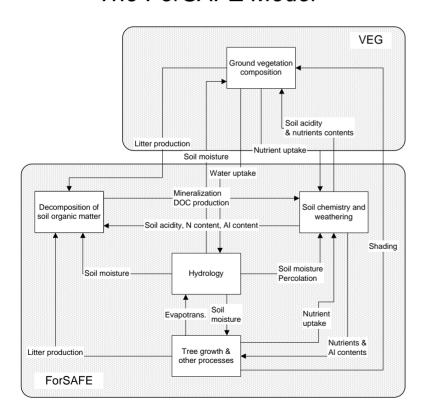
Comparison of critical loads estimated by the method of van Dobben et al. (2006) with those estimated by the Steady State Mass Balance

- Differences occur because:
 - The SMART model approach allows greater 'acceptable' N leaching than the DMB
 - Estimated N immobilisation is higher
- Compared to empirical critical loads, van Dobben approach gives similar range but no correlation for individual habitat types

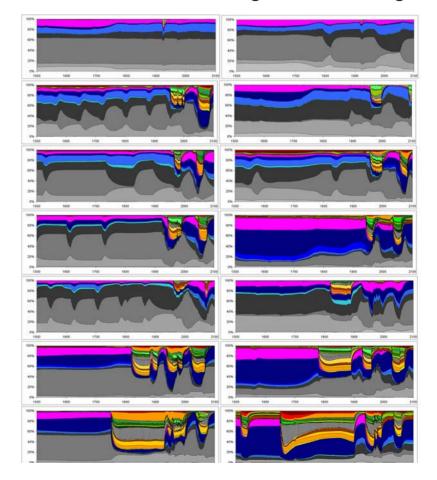


Calculating critical loads with dynamic models: 2. Sweden

The ForSAFE Model



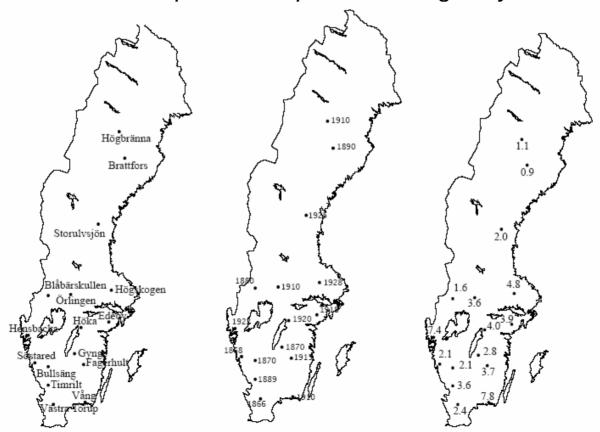
ForSAFE modelled vegetation change





Calculating critical loads with dynamic models: 2. Sweden

ForSAFE estimated critical loads based on the N deposition at which species composition changed by 5%





Calculating critical loads with dynamic models: 2. Sweden

ForSAFE critical loads and exceedances for individual sites

| Site | Time of vegetation response | Critical load deposition kg ha ⁻¹ yr ⁻¹ | Present deposition kg ha ⁻¹ yr ⁻¹ | Excess deposition kg ha ⁻¹ yr ⁻¹ | Required deposition reduction % |
|---------------|-----------------------------|---|---|--|--|
| Högbränna | 1910 | 1.1 | 1.5 | 0.4 | 27 |
| Brattfors | 1890 | 0.9 | 2.0 | 1.1 | 55 |
| Storulvsjön | 1925 | 2.0 | 3.5 | 1.5 | 43 |
| Högskogen | 1928 | 4.8 | 7.9 | 3.2 | 40 |
| Örlingen | 1910 | 3.6 | 8.5 | 3.9 | 52 |
| Edeby | 1918 | 3.9 | 7.8 | 3.9 | 50 |
| Blåbärskullen | 1880 | 1.6 | 8.5 | 6.9 | 81 |
| Höka | 1920 | 4.0 | 8.9 | 4.9 | 55 |
| Hensbacka | 1922 | 7.4 | 18.0 | 10.6 | 59 |
| Söstared | 1868 | 2.1 | 20.0 | 17.9 | 89 |
| Gynge | 1870 | 2.8 | 8.3 | 5.5 | 66 |
| Fagerhult | 1915 | 3.7 | 7.5 | 3.8 | 51 |
| Bullsäng | 1870 | 2.1 | 15.0 | 12.9 | 86 |
| Timrilt | 1889 | 3.6 | 23.0 | 19.4 | 84 |
| Vång | 1910 | 7.8 | 17.0 | 9.2 | 54 |
| Västra Torup | 1866 | 2.4 | 27.0 | 24.6 | 91 |



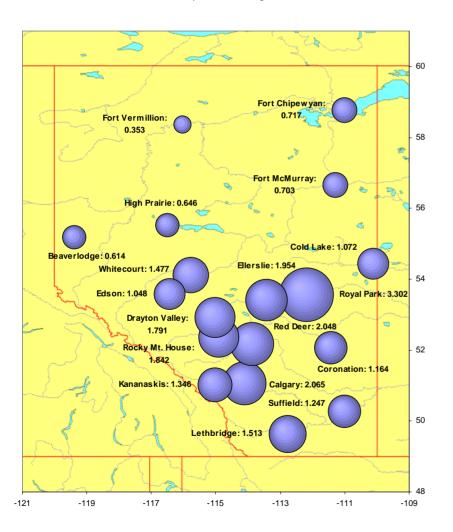
5. Critical Loads in Alberta

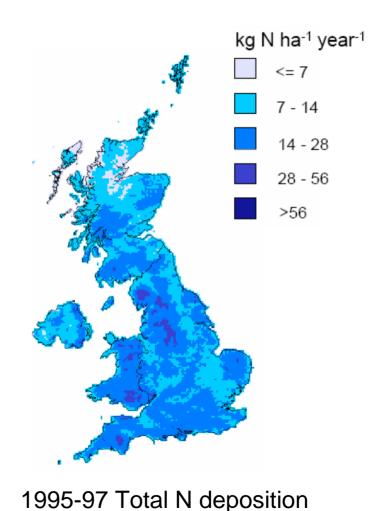




Alberta vs Europe: N deposition levels

Wet Deposition of Nitrogen





Alberta vs Europe: Acidity Critical Loads

- Acidity critical loads applied to both
- Methods appear fundamentally similar:

Net Acidifying Potential:

$$NAP = ([SO_4^{2-}] - [Ca^{2+} + Mg^{2+}])_{wet} + [NO_3]_{leached}$$

ForSust model: Steady state mass balance approach

- 95% protection level, similar chemical thresholds used
- Range of acidity critical loads (0.25 to 1.0 keq/ha/yr) similar to Europe, but with lower maximum values



Alberta vs Europe: Damage vs Recovery

- In Europe, critical loads are, or have been, exceeded across much of the area, so emphasis is on reduction of CL exceedance and modelling timescales of recovery
- In Alberta, critical loads haven't been exceeded anywhere, so emphasis is on avoiding damage

European Target Loads:

The target deposition required to achieve recovery by a specified date at a currently exceeded site:

'Have to do more'

Albertan Target Loads:

Somewhere between current deposition and the critical load (~90%).

'Factor of safety'



Alberta vs Europe: Eutrophication

- Critical loads for N as a nutrient have not yet been applied to Alberta
- Evidence from Europe is that ecosystems may be more sensitive to N deposition with regard to eutrophication than with regard to acidification
- One possibility is to adopt the critical loads for N as a nutrient developed in Europe
- But Albertan ecosystems and plant species differ significantly from those in Europe – need to ensure that sensitivity to N deposition is similar before applying European values.
- Ideally, a combination of experiments and linked soil-vegetation condition surveys are required to establish local species sensitivity to N deposition

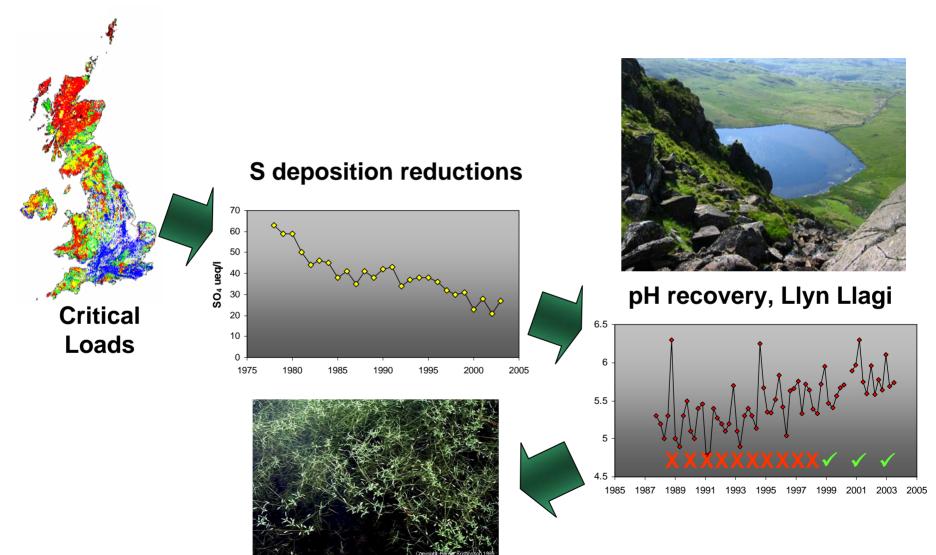


Conclusions

- Critical loads aren't perfect!
 - They do not consider timescales of change
 - They simplify complex ecosystem processes by which deposition impacts on environmental quality into 1 (or 2) numbers
 - Chemical criteria and damage thresholds are not always well defined or verified
 - Long-term sinks, particularly for N, are uncertain
 - They assume a threshold that might not really exist
- Dynamic models can address some of these limitations, but are unlikely to entirely replace critical loads
- And whatever their failings, critical loads have proven to be a highly effective means of translating science into policy, and take significant credit for the success of negotations to reduce acidifying emissions in Europe



Critical loads have worked...



Callitriche hamulata (water starwort)